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To cite this article before publication: Diling Liang *et al* 2024 *Environ. Res. Commun.* in press <https://doi.org/10.1088/2515-7620/ad37f3>

Manuscript version: Accepted Manuscript

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Monitoring Spatiotemporal Changes in Land Use/Land Cover and its Impacts on Ecosystem Services in Southern Zambia

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Abstract

Ecosystems play a vital role in human well-being, yet the widespread loss of ecosystem services due to human activities, including agricultural expansion and deforestation, remains a significant concern. Despite the wealth of research highlighting the importance of ecosystem conservation in Zambia, a critical gap exists in understanding the interplay between the conservation of ecosystem services and the socio-economic needs of local communities. This study presents a comparative analysis of ecosystem services in two distinct landscapes within southern Zambia: the protected area of Kafue National Park (KNP) and the agricultural landscape of Kalomo district between 2000-2020. Employing a combination of quantitative and qualitative methods, we evaluate the impacts of land/use and land cover (LULC) changes on selected ecosystem services, with a particular focus on carbon storage and the habitat quality of the trumpeter hornbill. The results of the comparison indicate that: (1) the Kalomo district has suffered from extensive land conversion, with forest changing to cropland, while KNP was well protected from encroachment, with forest area increasing over time; (2) carbon stocks and the habitat quality of trumpeter hornbills continually decreased in the Kalomo district but improved in KNP; (3) Kalomo district has suffered rapid environmental degradation due to an imbalance between economic development and environmental conservation, while strict enforcement in KNP has preserved ecosystems. The findings underscore the importance of integrated and inclusive land-use planning and natural resource governance for maintaining and enhancing ecosystem services in Zambia. To progress towards landscape management that is both

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3 31 sustainable and equitable, aligning with the objectives of the Global Biodiversity Framework, it
4 32 is proposed that a comprehensive approach be adopted in the region. This approach should
5 33 encompass a more thorough consideration of local livelihood requirements, as well as the wider
6 34 political-economic and social factors at play.
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12 36 **Keywords:** *Ecosystem services, InVEST model, LULC change, protected area, deforestation,*
13 37 *Zambia, Convention on Biological Diversity, Global Biodiversity Framework*
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3 59 **Declarations**

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5 60 **Ethical Approval**

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7 61 This study did not involve human or animal subjects. However, ethical principles were
8 62 considered throughout the study, and all research procedures were conducted in compliance with
9 63 ethical standards for scientific research.

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11 64 **Competing interests**

12
13 65 We confirm that we have no conflicts of interest associated with this publication. We confirm
14 66 that the manuscript has been read and approved for submission by all the named authors.

15
16 67 **Authors' contributions**

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18 68 All authors have contributed significantly to the study, as detailed below:

19
20 69 Diling Liang: Conceptualization, study design, data collection, data analysis and interpretation,
21 70 manuscript writing.

22
23 71 James Reed: Conceptualization, study design, data analysis and interpretation, visualization,
24 72 review and editing.

25
26 73 Sima Fakheran: Conceptualization, data analysis and interpretation, review and editing.

27
28 74 Kaala Moombe: Conceptualization, data analysis and interpretation, review and editing.

29
30 75 Freddie Siangulube: Conceptualization, data analysis and interpretation, review and editing.

31
32 76 Terry Sunderland: Conceptualization, study design, data analysis and interpretation,
33 77 visualization, review and editing.

34
35 78

36
37 79 **Funding**

38
39 80 This study was funded by the Center for International Forestry Research(CIFOR)

40
41 81 **Availability of data and materials**

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43 82 All data generated or analyzed during this study are included in this published article and its
44 83 supplementary information files. Additionally, the datasets used in this study are available from
45 84 the corresponding author upon reasonable request.

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86 1. Introduction

87 Ecosystem services (ES) are the benefits humans directly or indirectly obtain from
88 nature, categorized into four functional groups: provisioning, regulating, cultural, and supporting
89 services (Millennium Ecosystem Assessment [2005](#), Lanzas et al [2019](#), Liu et al [2010](#), Neugarten
90 et al [2018](#)). Human survival and quality of human life ultimately depend on this range of
91 ecosystem services (Summers et al [2018](#)). For instance, the availability of food, fresh water, and
92 shelter from ecosystems are the basic materials for human life. However, human activities have
93 drastically altered landscapes and ecosystems (Berihun et al [2021](#)), and approximately 60% of
94 ES are being degraded due to anthropogenic drivers such as land-use and climate change
95 (Millennium Ecosystem Assessment, [2005](#), Lanzas et al [2019](#)).

96 Land-use change is the primary direct driver of ecosystem degradation, thereby critically
97 affecting their ability to maintain the provision of goods and services to humanity (Li et al [2015](#),
98 Rai et al [2018](#)). Pressures resulting from increasing population, consumption, inequality,
99 urbanization, and globalization have led to increased demand for land, food, water, and other
100 natural resources. Such pressures have resulted in agricultural expansion, deforestation, land
101 degradation, and biodiversity loss, impacting human well-being by significantly altering the
102 provision of ecosystem services (Elmhagen et al [2015](#), Henderson and Loreau [2018](#), Reader et al
103 [2022](#)). Zambia's land-cover has undergone a series of complex changes during the past three
104 decades, largely due to social, political, and economic influences (Phiri et al [2019](#)), and these
105 changes have negatively impacted the provision of many ES. The ES framework is a
106 comprehensive approach employed for the management of ecosystems, with a notable impact on
107 the process of decision-making (Evers et al. [2018](#)). Consequently, there is an increasing
108 prevalence of policymakers seeking environmental sustainability assessments in relation to
109 development. However, the predominant focus of research on ecosystem services (ES) in Zambia
110 is around the use of case studies that examine a solitary landscape to evaluate and delineate ES,
111 commonly centred on provisioning services (Deuteronomy et al [2019](#), Van der Horst et al [2014](#),
112 Lehner et al [2021](#)).

113 In contrast to provisioning services, regulating services, cultural services, and supporting
114 services are particularly underestimated in Zambia. Furthermore, few studies have taken into
115 account the simultaneous consideration of conservation objectives and the social-economic

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3 116 demands of local communities. Notably, Zambia has just committed to the Kunming-Montreal
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5 117 Global Biodiversity Framework, which outlines a comprehensive set of measures designed to
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7 118 “enhance biodiversity and ecosystem functions and services, ecological integrity and
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9 119 connectivity,” while also emphasizing the “recognition and respect the rights of indigenous
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11 120 peoples and local communities.” This commitment exemplifies an increasing acknowledgement
12
13 121 of the significance of supportive services in advancing sustainable development and
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15 122 safeguarding the welfare of indigenous populations.

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17 123 In Zambia, a substantial portion of the land, approximately 40%, is designated as
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19 124 protected areas (PAs), comprising 20 national parks and 36 Game Management Areas (GMAs),
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21 125 all established with the primary aim of conserving biodiversity (Lindsey et al [2014](#), Hou-Jones et
22
23 126 al [2019](#), Lecina-Diaz et al [2019](#)). Zambian national parks are regarded as strict PAs where
24
25 127 human settlement is strictly prohibited and agricultural activities are forbidden (Lindsey et al
26
27 128 [2014](#)). Conversely, to promote socio-economic development, agriculture has been extensively
28
29 129 developed in Zambia. The country faces considerable deforestation rates, driven largely by
30
31 130 agricultural expansion (Richardson et al [2021](#)), resulting in an annual deforestation rate of
32
33 131 250,000 to 300,000 hectares per year (Phiri et al [2022](#)). Consequently, it becomes imperative to
34
35 132 comprehend the diverse impacts of various land management policies on LULC, ES, and local
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37 133 communities, and to propose integrated approaches to natural resource management at the
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39 134 landscape scale that effectively harmonize both conservation objectives and the needs of local
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41 135 communities in Zambia.

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43 136 To understand the influence of land management policies on ES, this study conducted an
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45 137 assessment and comparative analysis of the long-term (2000-2020) land-use dynamics on
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47 138 regulating services (carbon storage) and supporting services (habitat quality), in Kafue National Park
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49 139 and Kalomo district in Zambia. Kafue National Park represents a stringent PA, while Kalomo district
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51 140 epitomizes a well-established agricultural region within the Zambian landscape. The objectives of
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53 141 this study encompass the following: (1) evaluation of LULC changes within both landscapes, each
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55 142 subject to distinct land management policies; (2) quantification and spatial mapping of carbon
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57 143 storage and habitat quality for the Trumpeter hornbill employing ecological models; (3) comparative
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59 144 analysis of the land management policies that propel land-use alterations in the two contrasting
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145 landscapes; (4) formulation of recommendations for divergent landscape management strategies.
146 Through the pursuit of this research endeavour, our overarching objective is to enhance our

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3 147 comprehension of how to effectively harmonize the livelihood requirements of local
4 148 communities with the imperative task of conserving ES.
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8 149 **2. Methodology**

9 150 **2.1 Study sites**

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12 151 The study areas are Kafue National Park and Kalomo District (Figure 1). Both landscapes
13 152 are characterized by Miombo and Mopane woodlands. The Miombo woodlands of southern
14 153 Africa are one of the most significant dry forests in Africa (Rduch 2016, Phiri et al 2019) and are
15 154 significant sources of livelihood benefits and have critical functions in conserving biodiversity
16 155 and mitigating climate change (Chanda 2007, Moombe et al 2020).
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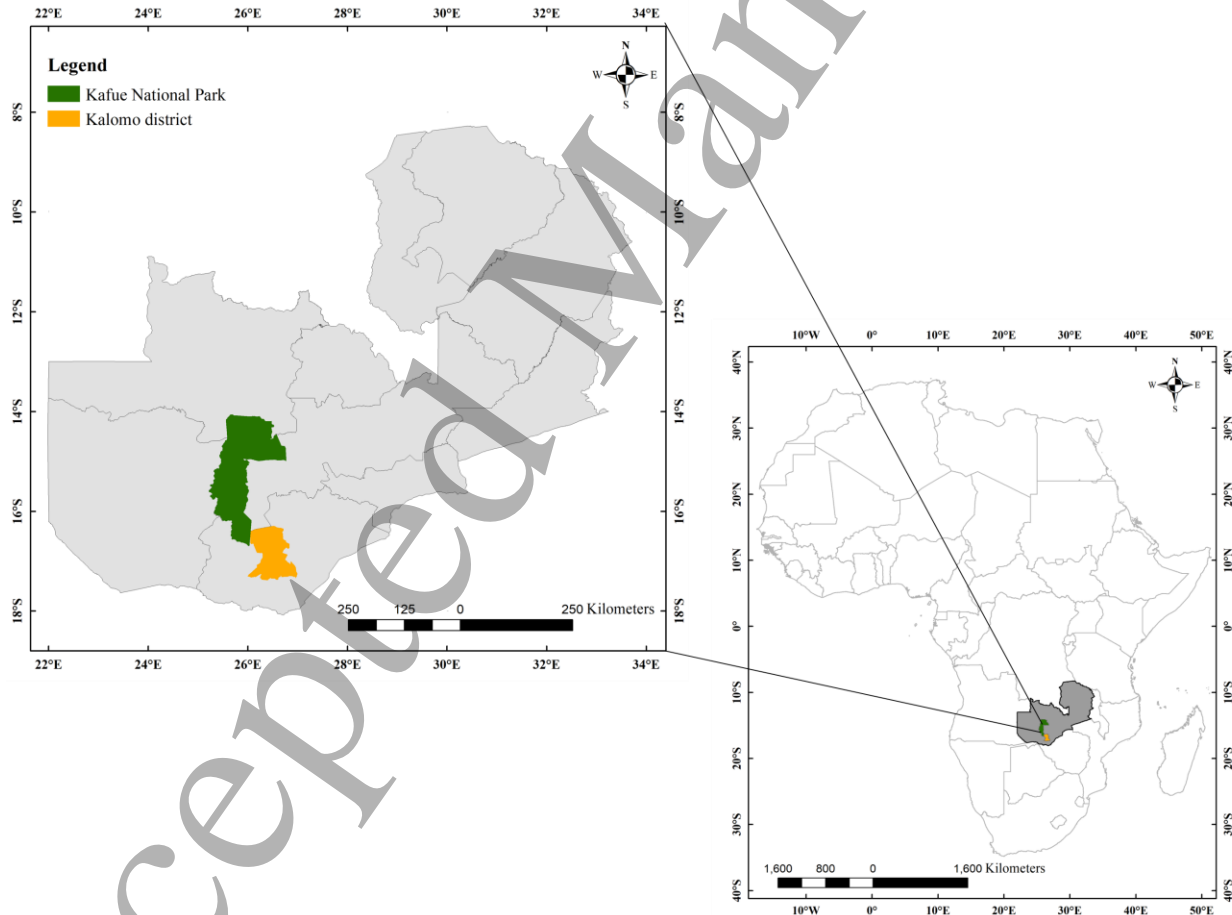


Figure 1. The locations of study areas in Zambia

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3 158 The KNP is located in the central area of southwestern Zambia. It is the oldest and largest
4 national park in Zambia, covering an area of 22,480 km² (Gula and Phiri [2020](#)). Nine Game
5 159 Management Areas (GMAs) border the KNP, and it is estimated that more than 174,796 people
6 160 live in the proximity of the KNP (Namukonde and Kachali [2015](#)). While KNP is a national asset
7 161 that brings rewards at the national level and is important for the conservation of unique
8 162 biodiversity, its stringent restrictions on access to natural resources have profound socio-
9 163 economic repercussions on the surrounding communities (Vezina et al [2020](#)). These
10 164 communities heavily rely on natural resources for their livelihoods, yet they are largely excluded
11 165 from the park to the extent that most people consider visiting the park to be illegal (Watson et al
12 166 [2014](#), Namukonde and Kachali [2015](#), Milupi et al [2021](#)).

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21 168 Kalomo district is located in the Southern province, covering 8,075 km² (Moombe et al
22 169 [2020](#)). Kalomo is a typical agricultural landscape in Zambia. It is referred to as the “Farmers’
23 170 Nest” because of the commercial, small to medium-scale livestock and crop (specifically maize)
24 171 farming enterprises (Bush [2014](#), Moombe et al [2020](#)). However, Kalomo district also ranks as
25 172 one of the economically poorest in the country (Sialubanje et al [2017](#)). Agriculture is
26 173 predominantly rain-fed and the primary economic resource for local people, accounting for
27 174 34.4% of local household income (Moombe et al [2020](#)). Maize is the primary staple crop in
28 175 Kalomo and occupies 55% of the cultivated area. Indigenous and hybrid maize varieties account
29 176 for 25% and 30% of the cultivated crops, respectively (Kalinda et al [2010](#)). In addition to crop
30 177 production activities, households keep livestock and cattle are the most important livestock
31 178 species owned by local farmers for various purposes (Kalinda et al [2010](#)). Moombe et al ([2020](#))
32 179 stated that in the Kalomo district as of 2020, the human population is 395,471 and the total
33 180 number of livestock is 411,765.

34 35 36 37 38 39 40 41 42 43 44 45 181 **2.2 Ecosystem Services Assessment**

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47 182 The Integrated Valuation of Ecosystem Services and Tradeoffs (InVEST) model is a
48 183 geospatial model to quantify and map ecosystem services (Keller et al [2015](#)). It operates on a
49 184 gridded map with an average annual time step, making it suitable for evaluating the effects of
50 185 land-use change on various ES (Tallis et al [2014](#)). The InVEST model includes different
51 186 modules, and carbon storage and habitat quality modules are used to estimate the status and
52 187 variation in carbon stock and habitat quality provided by the different LULC types (Maanan et al

188 [2019](#), Nematollahi et al [2020](#)). It is extensively utilized for quantifying ES due to its ability to
189 customize input data and settings, as well as its requirement for limited data for researched ES
190 (Grafius et al 2016, Daily et al [2009](#), Ochoa and Urbina-Cardona [2017](#)).

191 **2.2.1 Carbon storage**

192 Carbon sequestration is one of the key supporting services (Sintayehu [2018](#)). Preserving
193 carbon stocks is an important objective of the United Nations Framework Convention on Climate
194 Change (Soto-Navarro et al [2020](#)). The practical assessment of carbon stock could guide how
195 carbon targets can be incorporated into national policies and implemented in climate change
196 mitigation and adaptation (Soto-Navarro et al [2020](#)). The carbon storage module of InVEST can
197 estimate the spatial distribution of carbon stock across the study areas based on the simple
198 carbon cycle (Maanan et al [2019](#), Piyathilake et al [2022](#)). The total carbon storage of the
199 landscape is the sum of four different carbon pools assigned for each LULC type: (i)
200 aboveground biomass, which includes all living plant materials; (ii) belowground biomass,
201 which comprises the living roots systems; (iii) soil organic matter; and (iv) dead organic material
202 (Sialubanje et al [2017](#), Dietz [2021](#), Duarte et al [2016](#), Piyathilake et al [2021](#)).

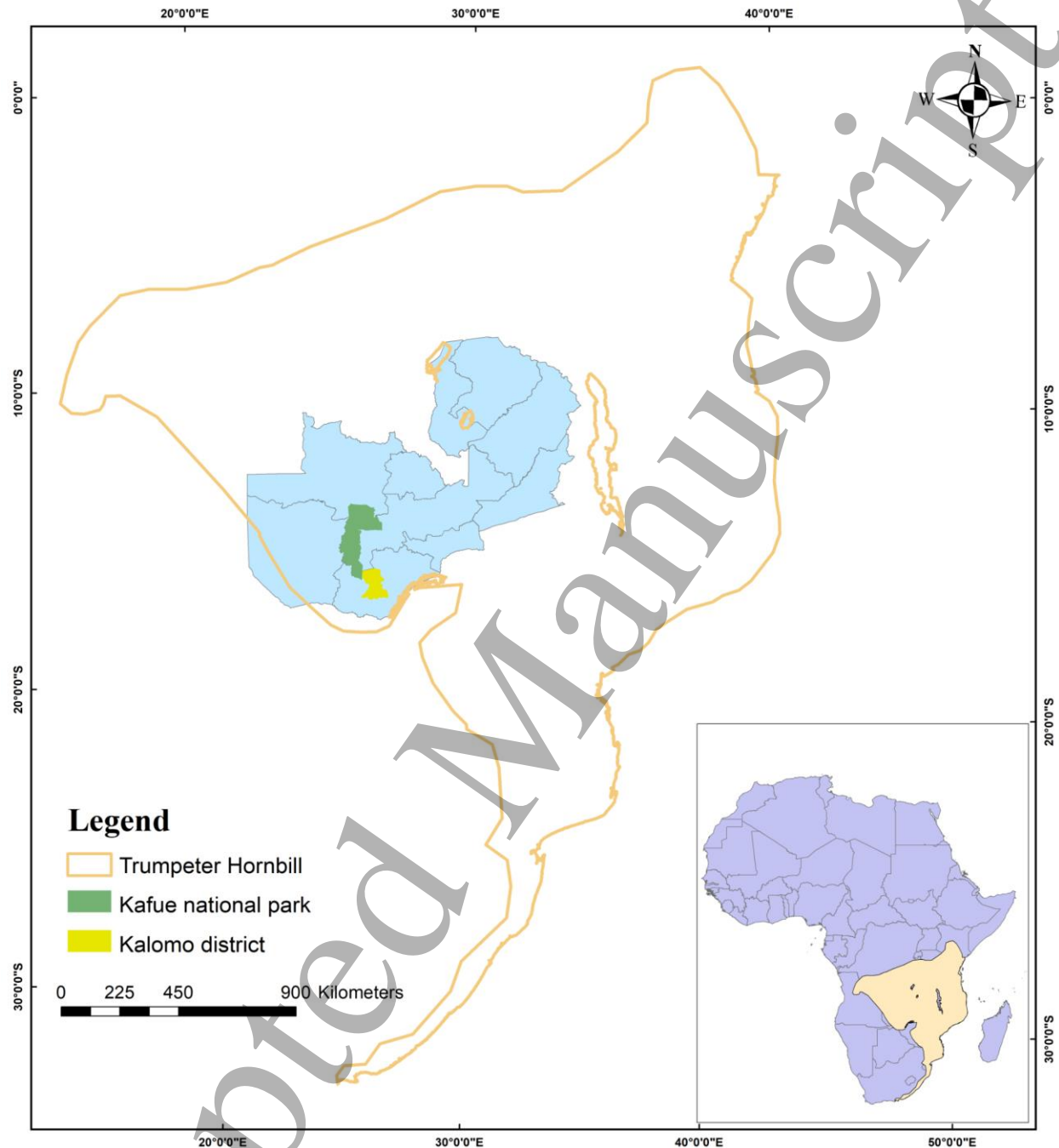
203 **2.2.2 Habitat quality**

204 Birds serve as valuable tools for biodiversity monitoring in forest ecosystems due to their
205 sensitivity to environmental changes and their ease of identification (Copper et al [2020](#)). This
206 unique combination of attributes makes monitoring programs not only feasible but also
207 accessible to both scientists and the general public (Copper et al [2020](#), Wade et al [2013](#)). In our
208 study, we have chosen the Trumpeter hornbill (*Bycanistes buccinator*) as the focal species due to
209 its distribution encompassing both the KNP and the Kalomo district (Figure [2](#)). Unlike the
210 Ground-Hornbill (*Bucorvus leadbeateri*), with limited habitat preference in KNP and listed as
211 Vulnerable (Gula and Phiri [2020](#)), the Trumpeter hornbill has been a Least Concern species on
212 the IUCN Red List since 1984 (BirdLife International [2018](#)). However, trumpeter hornbills are
213 considered large, frugivorous birds, facilitating the functional connectivity of fragmented
214 landscapes. This is attributed to their remarkable ability to disperse seeds over considerable
215 distances and among suitable habitat patches (Mueller et al [2014](#), Lenz et al [2015](#)). Such actions
216 support gene flow, range expansion, and natural forest regeneration, all of which are crucial for
217 biodiversity conservation (Mueller et al [2014](#), Lenz et al [2015](#)).

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3 218 The miombo woodlands in Zambia have a high diversity of plants, animals, and avifauna,
4 219 while also serving vital ecological roles (Malunga et al [2021](#)). However, the miombo woodlands
5 220 are heavily fragmented due to anthropogenic interference, with only a few natural forest patches
6 221 remaining. Safeguarding the trumpeter hornbill within Zambia's miombo woodlands is crucial to
7 222 maintaining biodiversity and landscape connectivity amid escalating habitat fragmentation. Thus,
8 223 it is important to monitor and evaluate the status of the trumpeter hornbill to ensure that it
9 224 remains protected.

10 225 The habitat quality module of InVEST combined “information on LULC and the threats
11 226 to biodiversity to produce habitat maps” (Tallis et al [2014](#)). This module used habitat quality and
12 227 rarity as proxies to represent the biodiversity of a landscape, and it also estimated the extent of
13 228 habitat and vegetation type throughout the landscape, as well as their level of degradation (Tallis
14 229 et al [2014](#)). This module was based on the hypothesis that “areas with higher habitat quality
15 230 supported higher richness of native species, and that decrease of habitat extent and quality led to
16 231 a decline in species persistence” (Terrado et al [2016](#)). This module was used to assess how the
17 232 study areas provide suitable habitat for Trumpeter hornbills based on available data. The value of
18 233 habitat quality is assigned to LULC types to show habitat suitability. The range of the value is
19 234 between 0 to 1, where 1 indicates the highest habitat quality, while 0 means unsuitable habitat for
20 235 certain species (Terrado et al [2016](#), Chu et al [2018](#)). To visually display the habitat quality
21 236 according to the output values of the habitat quality model, the equal interval breakpoint method
22 237 in ArcGIS was applied to assign grades to three different habitat quality groups (Wang et al
23 238 [2022b](#)).

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240
241 **Figure 2. The distribution of trumpeter hornbill**

242 **2.3 Data Resources**

243 The types of data used in this study mainly include land cover/land use data (used to
244 assess land use change), satellite images (used to validate the accuracy of land-use

245 classification), carbon stock parameters (used for carbon storage assessment), and other data.
 246 The specific data resources and their descriptions are shown in Table 1. All data is open-access.

247 **Table 1. Data sources and descriptions**

Data name	Data description	Data resource
Land use/Land cover	30m-resolution land-cover datasets in 2000 and 2010	https://rcmrd.africageoport.com/
Land use/Land cover	30m-resolution land-cover maps in 2020	Supervised Classification in the Environment for Visualizing Images (ENVI)
Satellite images	Landsat-5 and Landsat-8 images	https://www.usgs.gov/
Carbon stock parameters	Carbon storage data for different land use/land cover types	Dietz (2021); Day et al (2014); Forest Reference Emission Level (2021); Gumbo et al (2018); IPCC 2006 report; Sialubanje et al (2017); Piyathilake et al (2022)
Habitat quality parameters of Trumpeter Hornbill	Threat factors; sensitivity of each land-use class; maximum distance	Expert knowledge
Road data	Main roads in Zambia	http://riskprofilesundrr.org/
Administrative boundary shapefile	Obtaining study area shapefile by cropping	https://data.grid3.org/
Trumpeter Hornbill	Trumpeter Hornbill's distribution map	https://datazone.birdlife.org/

248 In addition to using the above data, the land cover classification of the study area referred to the *Guidelines for National*
 249 *Greenhouse Gas Inventories (IPCC) in 2006*.

250

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3. Results

3.1 Land use and land cover change

The results of the study areas are presented in Figures 3 and 4, and the area of land cover and land-cover change in the study areas are summarized in Table 2. In KNP, there was no significant land-use conversion observed from LULC maps. Between 2000 and 2020, the land-use cover was dominated by grassland, and it was mainly distributed in the middle and northern parts of the study area. The area of grassland decreased from 59% (13,245 km²) in 2000 to 49% (10,832 km²) in 2010, but subsequently increased to 53% (11,868 km²) of the total area in 2020. The forest was the second dominant land-use type, and the area of the forest increased from 38% (8484 km²) in 2000 to 45% (10,070 km²) in 2020. The area of water bodies was relatively stable, and their area decreased from 1.8% (398 km²) in 2000 to 1.5% (331 km²) in 2020. Both water bodies and wetlands decreased from 2000 to 2020, losing 67 km² and 66 km², respectively.

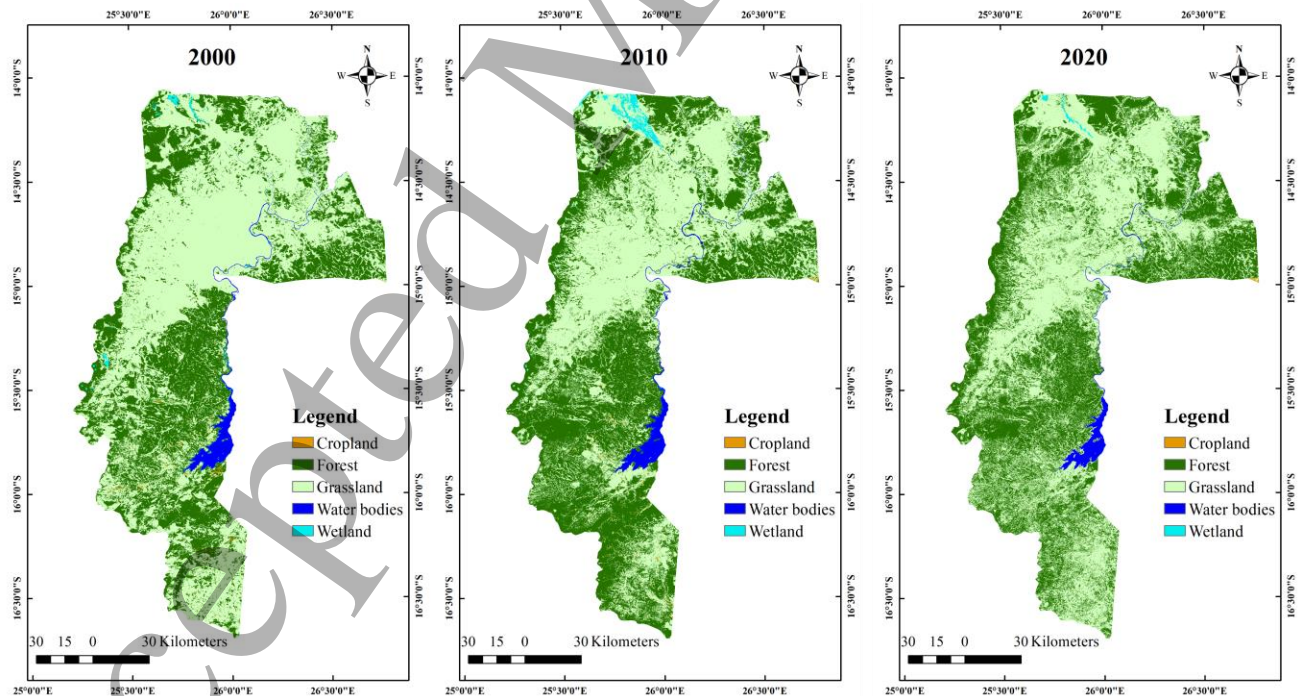


Figure 3. Spatial and temporal LULC in KNP in 2000, 2010 and 2020

268 **Table 2. LULC changes in 2000, 2010, and 2020 for KNP and Kalomo**

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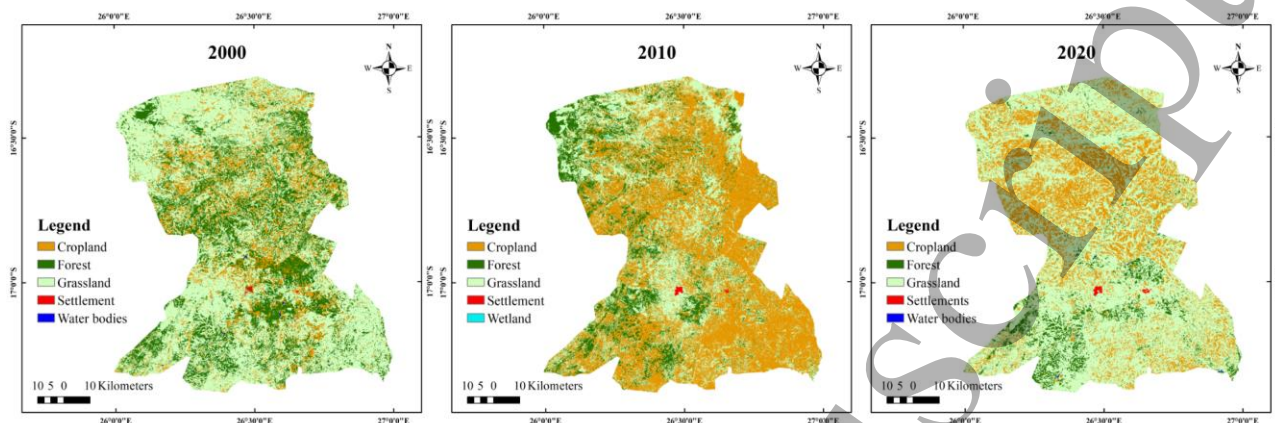
Areas	LULC Area (km ²)	2000	2010	2020	Changes(2000-2020)
KNP	Cropland	80.86 (0.36) *	76.03 (0.34)	4.56 (0.02)	-76.30 (-0.34)
	Forest	8484.50 (38.03)	10767.03 (48.26)	10070.34 (45.14)	1585.83(7.11)
	Grassland	13244.92 (59.37)	10832.22 (48.55)	11867.85 (53.20)	-1377.07 (-6.17)
	Water bodies	398.01 (1.78)	409.32 (1.83)	331.22 (1.48)	-66.79 (-0.30)
	Wetland	100.24 (0.45)	223.86 (1.00)	34.41 (0.15)	-65.83 (-0.29)
Kalomo District	Cropland	1263.38 (15.64)	4058.34 (50.26)	2158.15 (26.73)	894.76 (11.08)
	Forest	2115.29 (26.19)	1313.73 (16.27)	544.39 (6.74)	-1570.90 (-19.45)
	Grassland	4684.10 (58.01)	2685.28 (33.25)	5355.42 (66.33)	671.32 (8.32)
	Settlement	4.55 (0.06)	8.88 (0.11)	10.90 (0.14)	6.35 (0.08)
	Water bodies	7.96 (0.10)	0.00 (0)	5.53 (0.07)	-2.43 (-0.03)
	Wetland	0.00 (0)	9.12 (0.11)	0.00 (0)	0.00 (0)

270 Note: (0.36) *: Percentage of each LULC type (%), others as same. The total area of the KNP
 271 and Kalomo district is 22,308. 85 km² and 8,075.32 km², respectively.

272

273 In the Kalomo district, there was a significant land-use transition between 2000 and 2020.
 274 Generally, the pattern of land use transition was characterized by the change from forest to
 275 cropland from 2000 to 2020. The forest decreased from 26% (2115 km²) in 2000 to 16% (1314
 276 km²) in 2010, and then continued to decline to 7% (544 km²) in 2020. However, the cropland
 277 sharply increased from 15% (1263 km²) in 2000 to 50% (4058 km²) in 2010, and then
 278 decreased to 27% (2158 km²) in 2020. The grassland was the dominant land-use type in 2000,
 279 accounting for 58% (4684 km²). However, it decreased to 33% (2685 km²) in 2010 but
 280 increased to 66% (5355 km²) in 2020. The settlement steadily increased from 0.06% (5 km²) in
 281 2000 to 0.14% (11 km²). The rest of the other land-use types, including water bodies and

282 wetland, only occupied a small area, and the percentage of these LULC types is less than 0.1%.



283
284 **Figure 4. Spatial and temporal LULC in Kalomo in 2000, 2010 and 2020**

286 3.2 Assessment of Ecosystem Services

287 3.2.1 Carbon storage

288 **KNP.** The spatial changes of carbon storage in KNP are spatially shown in Figure 5. The
289 maps show that the range of carbon stock in each grid cell is from 0 to 12.56 t/C. During 2000-
290 2020, the carbon storage of KNP slightly changed and exhibited overall growth. Total carbon
291 storage values were 189, 209 and 204 million t/C in 2000, 2010, and 2020, respectively. The
292 carbon value changed with the exchange of LULC classes. The corresponding average carbon
293 densities were 8.24, 8.43 and 7.66 tons per grid cell in 2000, 2010, and 2020, respectively. The
294 carbon storage was mainly distributed in the southern and eastern areas because forests occupied
295 these areas and exhibited an increase in carbon storage from 2000 to 2010, then decreased from
296 2010 to 2020. However, less carbon storage appeared in the central-northern area because
297 grassland has a lower ability to store carbon. Carbon stored in this area increased gradually from
298 2000 to 2020 because of decreasing grassland. There was no carbon storage in Itzhi-Tezhi Lake
299 and Kafue River because the carbon stored in such water bodies is negligible.

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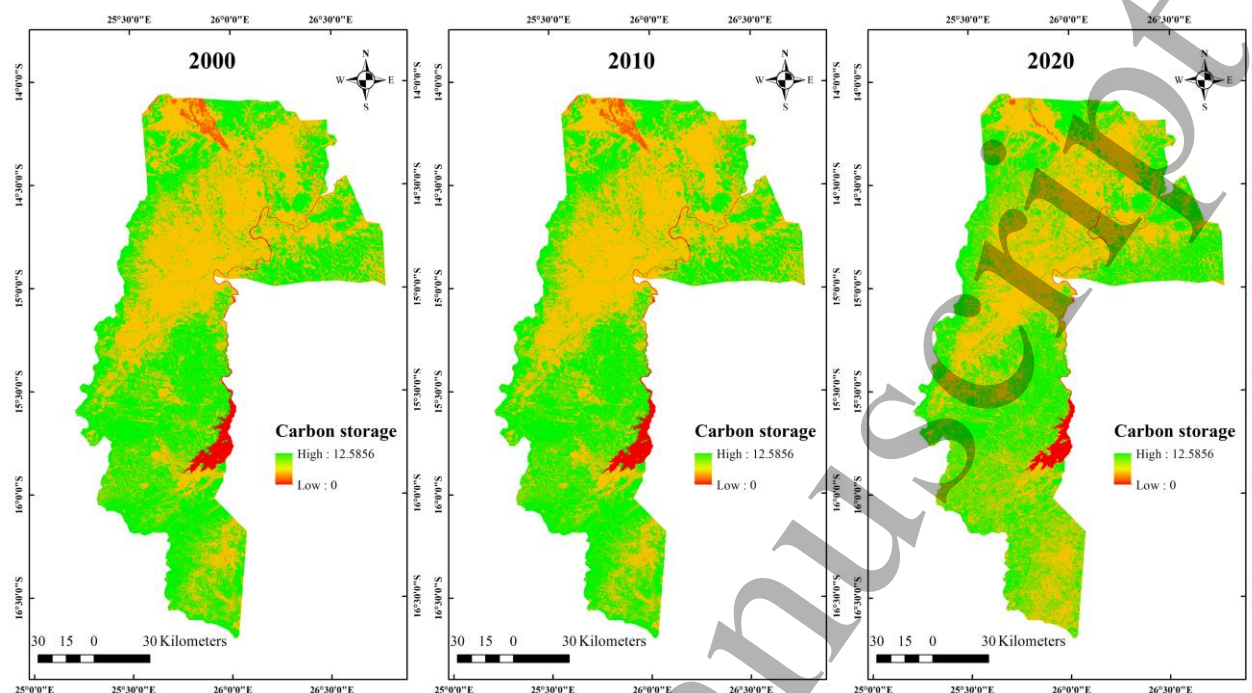
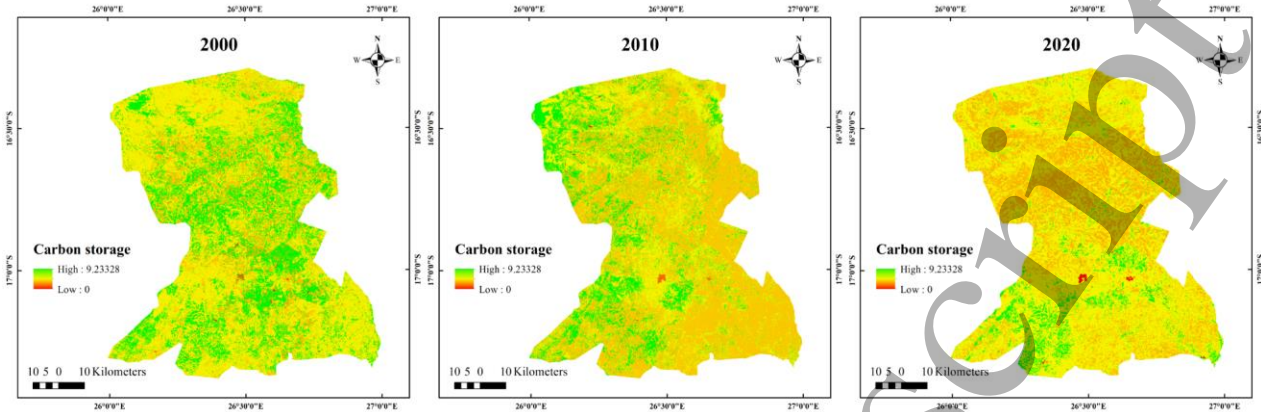


Figure 5. Spatial distribution of carbon storage in KNP

Kalomo. The change in carbon storage in Kalomo is spatially shown in Figure 6. The carbon storage of the Kalomo district continually decreased from 2000 to 2020. The average carbon densities were 5.8, 5.0, and 4.8 tons per grid cell in 2000, 2010, and 2020, respectively. The total carbon storage values were 52.04, 44.94, and 43.31 million t/C, respectively. Reduced carbon storage was caused by severe land conversion, whereby forest was converted into cropland. Over the past twenty years, there have been about 9 million tons of carbon loss due to agricultural expansion and associated forest loss. Regarding spatial distributions, carbon storage mainly decreased across the study area from 2000 to 2010, except for the northwestern areas because agriculture expanded from the center to the periphery. Then carbon continually decreased from 2010 to 2020, and limited carbon remained in the center and southwestern area in 2020 because the fallow areas had relatively low vegetation coverage.



315
316 **Figure 6. Spatial distribution of carbon storage in Kalomo district**

317 3.2.2 Habitat quality

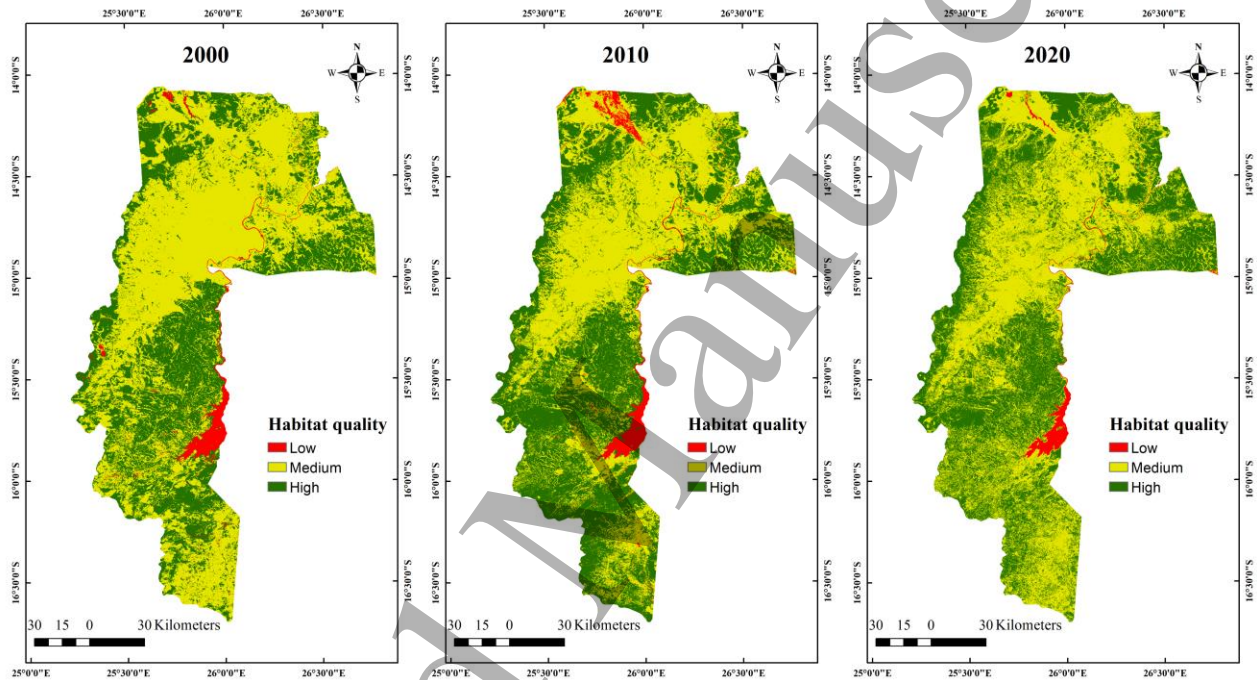
318 To visually represent the output values of the habitat quality model, the equal interval
319 breakpoint method was applied (Wang et al [2022b](#)). This method was used to assign grades to
320 three habitat quality groups, which were designated as low, medium, and high, and these classes
321 represent poor habitat quality, medium habitat quality, and high habitat quality, respectively
322 (Table [3](#)). Spatial distributions of habitat quality maps are shown in Figures [7](#) and [8](#), and the
323 statistical analysis of the changes in habitat quality is presented in Table [4](#). The area in green
324 shows high habitat quality, while the area in red shows poor habitat quality.

325 **Table 3. Classification values of trumpeter hornbill's habitat quality in study areas**

Habitat quality grade	Range of values	Description
Low	0 – 0.33	Poor habitat quality
Medium	0.33 – 0.67	Medium habitat quality
High	0.67 - 1	High habitat quality

326
327 **KNP.** Overall, habitat quality for trumpeter hornbill in KNP has improved from 2000 to
328 2020, with an increase of 1586 km² (7.11%) in high habitat quality, but a decrease of 209
329 km² (0.94%) and 1377 km² (6.17%) in poor and medium habitats, specifically (Table [4](#)). The
330 overall habitat grade increased from 0.59 in 2000 to 0.65 in 2010 but slightly decreased to 0.64
331 in 2020. From Figure [7](#), we can see that medium and high grades dominated the habitat quality
332 of KNP because grasslands and forests occupied 97% of the whole area. The area with high
333 habitat quality grade was mainly distributed in the south and northeastern regions, where there is

334 greater coverage of forest. These areas increased from 8484 km² (38.03%) to 10767
 335 km² (48.27%) in 2010, but slightly decreased to 10070 km² (45.14%). The areas with medium
 336 habitat quality were mainly distributed in the central region, decreasing from 13245
 337 km² (59.37%) in 2000 to 10832 km² (48.56%) in 2010, but increasing to 11868 km² (53.20%)
 338 in 2020. The areas with poor habitat quality increased from 579 km² (2.59%) in 2000 to 709
 339 km² (3.18%) in 2010, but decreased to 370.19 km² (1.66%) in 2020.



340
 341 **Figure 7. Spatial distribution pattern of trumpeter hornbill's habitat quality in KNP, in 2000, 2010,**
 342 **and 2020**

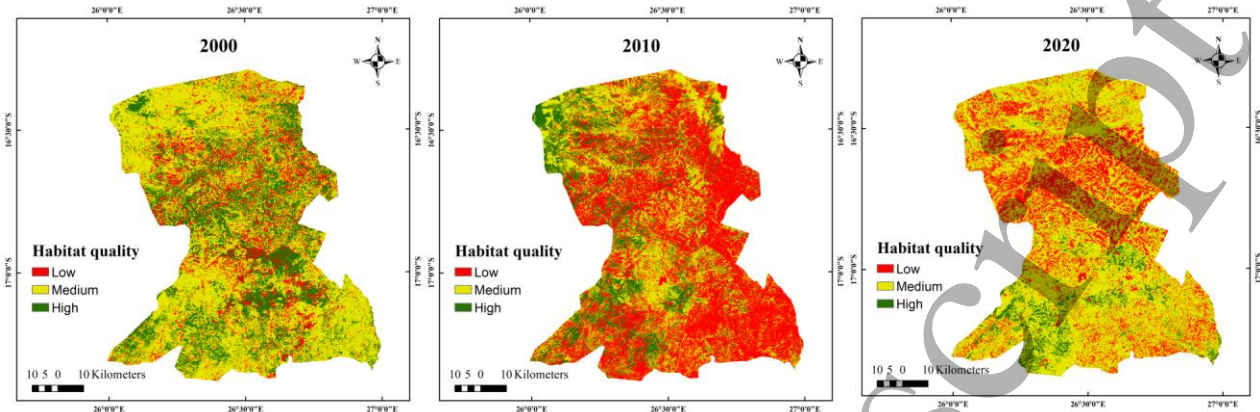
343 **Table 4. Area and percentage of trumpeter hornbill's habitat quality in KNP and Kalomo (km²,%)**

Areas	Classes	2000	2010	2020	Changes (2000-2020)
KNP	Low	578.83 (2.59) *	709.07 (3.18)	370.19 (1.66)	-208.64 (-0.94)
	Medium	13245.31 (59.37)	10832.21 (48.56)	11867.94 (53.20)	-1377.37(-6.17)
	High	8484.36 (38.03)	10767.21 (48.27)	10070.34 (45.14)	1585.98 (7.11)
Kalomo District	Low	1277.71 (15.82)	4076.92 (50.49)	2175.36 (26.94)	897.65 (11.12)
	Medium	4683.71 (58.00)	2684.47 (33.24)	5355.42 (66.32)	671.71 (8.32)
	High	2113.80 (26.18)	1313.63 (16.27)	544.38 (6.74)	-1569.42 (-19.44)

344 Note: (2.59) *: Percentage of each habitat quality class (%), others as same. The total area of the
 345 KNP and Kalomo district is 22,308. 85 km² and 8,075.32 km², respectively.

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3 347 **Kalomo.** Overall, the habitat quality for the trumpeter hornbill in the Kalomo district was
4
5 348 poor and dominated by low and medium grades. The average habitat quality scores were 0.44,
6
7 349 0.26, and 0.27 in 2000, 2010, and 2020, respectively. These values indicated that the trumpeter
8
9 350 hornbill's habitat degraded between 2000 and 2020. The three groups of habitat quality in
10
11 351 Kalomo are shown in Table 4. A total of 1569 km² of high habitat quality were lost from 2000 to
12
13 352 2020, representing 19.44% of the total area. From the spatial distribution maps (Figure 8), the
14
15 353 habitat quality distribution was highly consistent with the distribution characteristics of land use
16
17 354 types. High habitat quality was mainly concentrated in the central and southwestern areas in
18
19 355 2000, accounting for 2114 km² (26.18%), because these were areas with a concentrated
20
21 356 distribution of forest. However, poor habitat quality was distributed in the central and
22
23 357 northeastern regions, which were occupied by agriculture, accounting for 1278 km² (15.82%).
24
25 358 Medium habitat quality was evenly distributed in the rest of the areas, with 4684km², accounting
26
27 359 for 58.00% of the total area. From 2000 to 2010, the habitat quality degraded from east to west,
28
29 360 and most areas of high habitat quality were replaced by poor habitat quality. The limited high
30
31 361 habitat quality was distributed in the northwestern and southwestern regions, with 1314
32
33 362 km², only accounting for 16.27 %. While poor habitat quality dominated in 2010, occupying
34
35 363 half of the area (50.49%) with 4077 km². These changes were caused by agricultural expansion
36
37 364 between 2000 and 2010. From 2010 to 2020, the area of high habitat quality continually
38
39 365 decreased, and only a small portion of high habitat quality located in the southwestern regions
40
41 366 was retained, with 544km² (6.74%). The poor habitat quality decreased from 4076 km²
42
43 367 (50.49%) in 2010 to 2175 km² (26.94%) in 2020, but the medium habitat quality increased from
44
45 368 2684 km² (33.24%) in 2010 to 5355 km² (66.32%).
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371 **Figure 8. Spatial distribution pattern of trumpeter hornbill's habitat quality in Kalomo, in 2000,**
 372 **2010, 2020**

373 4. Discussion

374 This study pioneers the examination of the impacts of diverse land cover changes on ES
 375 in southern Zambia over a period of 20 years. It introduced an effective methodology for
 376 monitoring land-use changes' influence on carbon storage and habitat quality in two contrasting
 377 environments: the agricultural landscape of Kalomo district and the protected area of KNP. The
 378 results indicated that KNP has effectively maintained its forest and grassland ecosystems by
 379 preventing substantial land-use transitions between 2000 and 2020. However, Kalomo district
 380 has undergone a rapid and dramatic transformation, with forests being converted into cropland
 381 due to the surging demands of agriculture, accompanied by shifts in grassland and settlement
 382 patterns. These transitions were primarily influenced by disparate management strategies. While
 383 KNP has remained under stringent protection measures to deter human interference, agricultural
 384 activities have witnessed substantial growth in the Kalomo region over the past two decades. The
 385 study also found that carbon storage and overall habitat quality, as evidenced by the distribution
 386 of the trumpeter hornbill, have been enhanced in KNP but degraded in Kalomo district. These
 387 results indicate that LULC changes impacted carbon stock and habitat quality, and this
 388 conclusion is supported by further evidence (Solomon et al [2018](#), Fusco et al [2021](#)).

389 4.1 Analysis of land-use change

390 The results of this study indicate that KNP avoided deforestation between 2000 and 2020,
 391 and forest areas significantly increased from 38% (8484 km²) in 2000 to 45% (10070 km²) in

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2
3 392 2020. This finding differs from common conclusions about the effectiveness of PAs in Africa.
4
5 393 Some scholars state that establishing PAs cannot prevent habitat loss but could reduce forest loss
6
7 394 to mitigate such loss (Cazalis et al [2020](#), Riggio et al [2019](#)), Rosa et al [2018](#), Bowker et al [2017](#)).
8
9 395 In contrast to these studies, our results show opposing trends of land cover change at the single-
10
11 396 park scale and support that land transition could be prevented and vegetation cover could
12
13 397 increase due to the effects of conservation implementation. The cropland in KNP represented
14
15 398 less than 1%, decreasing from 0.34% (76 km²) in 2000 to 0.02% (4.56 km²) in 2020. This
16
17 399 reduction can be attributed to restricted cultivation activities within the park, emphasizing its
18
19 400 commitment to conservation, as noted by Mwima in ([2001](#)).

20
21 401 However, land-use change in the Kalomo has undergone a typical trajectory from forest
22
23 402 conversion to cropland for commercial crops. The forest area continually decreased from 26%
24
25 403 (2115 km²) in 2000 to 7% (544 km²) in 2020, while cropland sharply increased from 15%
26
27 404 (1263 km²) in 2000 to 27% (2158 km²) in 2020. These changes can be attributed to various
28
29 405 factors, including leakage from KNP and the escalating demand for food and charcoal from
30
31 406 neighbouring provinces. The establishment of KNP resulted in the forced relocation of at least
32
33 407 five chiefdoms, and these communities heavily rely on natural resources for their livelihoods
34
35 408 (Namukonde and Kachali [2015](#)). Such displacement has led to intensified pressure on resources
36
37 409 outside the park. Studies have demonstrated significant LULC change in GMAs and found that a
38
39 410 total of 125,108 ha of forest being converted into cropland (Dietz [2021](#)). Kalomo district is in
40
41 411 proximity to KNP and its nine GMAs, so it's plausible to infer that deforestation in Kalomo is
42
43 412 closely linked to leakage from the park. Furthermore, the rampant deforestation in Kalomo
44
45 413 district is fueled by agricultural expansion and charcoal production. Policy initiatives such as the
46
47 414 establishment of farm blocks in Zambia, including Kalomo district, have stimulated investment
48
49 415 in agriculture, exacerbating land conversion (Chilombo [2021](#)). Urbanization, a prominent trend
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51 416 in Zambia, has heightened the demand for charcoal, particularly in densely populated areas like
52
53 417 Lusaka and Copperbelt, which heavily rely on imports from other provinces (USAID [2017](#)). This
54
55 418 is a general trend elsewhere due to population dynamics, rising inequality, migration, agricultural
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57 419 commodity expansion, legal and illegal extraction of natural resources, urbanization, and poor
58
59 420 and overlapping or incompatible governance structures (Moombe et al [2020](#), Reed et al [2022](#)).
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2
3 421 There are no settlements within KNP, but the area of settlements in Kalomo district has
4
5 422 consistently increased from 2000 to 2020. The primary factors behind the relocation of
6
7 423 communities from KNP for conservation purposes were identified by Mwima in (2001).
8
9 424 However, it is worth noting that the population of the Kalomo is experiencing a growth rate of
10
11 425 4.4% (CSO, 2010). The water bodies and wetlands in both regions exhibited a persistent decline
12
13 426 from 2000 to 2020. Specifically, in KNP, the area of water bodies and wetlands decreased by
14
15 427 67 km² and 66 km², respectively. The wetlands of the Kalomo district disappeared in 2020. The
16
17 428 primary cause of these alterations can be traced to the impact of drought. Climate change has
18
19 429 caused many extreme environmental events in Zambia, including more frequent and intense
20
21 430 seasonal droughts, increased valley temperature and extended periods of drought (Rosen et al
22
23 431 2021). Musonda et al (2020) stated that the drought trend in Zambia significantly increased
24
25 432 between 1981 and 2017, and twice severe droughts in 2005-2006 and 2015-2016, leading to
26
27 433 serious concern for agricultural and hydrological sectors in drought-prone areas of southern
28
29 434 Zambia (Musonda et al 2020). In addition, Kalomo district is situated in one of the most arid
30
31 435 areas of Zambia, characterized by ephemeral and non-perennial streams and rivers that rapidly
32
33 436 diminish after the rainy season (Republic-of-Zambia 2021). Deforestation in Kalomo has
34
35 437 resulted in the loss of vegetation cover along riparian zones, accelerating the depletion of water
36
37 438 sources (Republic-of-Zambia 2021). The district, predominantly inhabited by smallholder
38
39 439 farmers, witnesses unsustainable agricultural practices, including direct extraction of water from
40
41 440 rivers and streams for agriculture and livestock. These practices exacerbate the drying process of
42
43 441 these vital water sources (Upla et al 2022).

42 442 **4.2 Analysis of Carbon Storage Change**

43 443 Carbon storage increased in KNP but declined in Kalomo from 2000–2020. The
44
45 444 dynamics of LULC crucially impact ecosystem service provision (Rai et al 2018), with
46
47 445 forestland playing a significant role in carbon sequestration and storage in the miombo
48
49 446 woodlands (Pelletier et al 2018). Therefore, changes in forest land significantly impact carbon
50
51 447 storage. In KNP, the growth of forest land significantly enhanced carbon storage. The increase in
52
53 448 forest cover is the key to improving carbon sequestration due to trees role as carbon sinks (Nunes
54
55 449 et al 2019). This finding is consistent with previous studies that showed that increased vegetation
56
57 450 cover in PAs is effective at preventing carbon loss (Lobora et al 2017). In Kalomo district, the

1
2
3 451 decrease in the forest during 2000–2020 was attributed to deforestation caused by agricultural
4 452 expansion and charcoal production driven by rapid urbanization and as a response to
5
6 453 environmental shocks such as droughts, both locally and nationally. Between 2000 and 2010,
7
8 454 many miombo forests were cleared for agriculture and charcoal production, contributing to
9
10 455 carbon loss. This is also consistent with previous findings. Williams et al ([2008](#)) stated that the
11
12 456 clearance for agriculture could reduce the loss of stem wood carbon stocks. Also, Bulusu et al
13
14 457 ([2021](#)) and Gumbo et al ([2018](#)) found that deforestation in miombo woodlands was driven by
15
16 458 land clearance, and grassland and forest were cleared for agriculture and wood extraction for
17
18 459 energy. These land transitions led to a decrease in carbon stored across miombo woodlands (Jew
19
20 460 et al [2016](#)).

21 461 **4.3 Analysis of Habitat Quality Change**

22
23
24 462 The habitat quality of trumpeter hornbills has been enhanced in KNP but degraded in
25
26 463 Kalomo district. Trumpeter hornbill is a forest-dependent species that feeds on fruits (Chibesa
27
28 464 and Downs [2017](#)). Grassland can also provide food resources for the trumpeter hornbill, as this
29
30 465 species adds other food resources to meet their food requirements, such as insects and small
31
32 466 reptiles (Lenz et al [2015](#)). Therefore, habitat quality change is related to forest and grassland
33
34 467 changes. In KNP, the predominant land-use change was an increase in forests during 2000-2020,
35
36 468 with the increased forest cover providing more habitat and food for trumpeter hornbills,
37
38 469 effectively improving the habitat quality of this species. This finding is consistent with Cazalis et
39
40 470 al ([2020](#)), who showed that PAs were effective at conserving forest-dependent bird species. In
41
42 471 the Kalomo district, the Kalomo Hills Local Forest Reserve was destroyed by agricultural
43
44 472 encroachment following settlement (Moombe et al [2020](#), Mbanga et al [2021](#)). The settlement of
45
46 473 more than 12,700 farmers in the reserve has led to an increased demand for agricultural
47
48 474 production, resulting in the conversion of forests and grasslands into productive land for crops,
49
50 475 particularly maize, to meet household income needs through sales (Mbanga et al [2021](#)), leading
51
52 476 to a decline in the habitat quality of trumpeter hornbill. Increased charcoal production to satisfy
53
54 477 urban demand has further and significantly contributed to forest loss.

55 478 **4.4 Analysis of the Impact of Land Management Policies**

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2
3 479 KNP is a national asset that brings benefits at the national level and is important for the
4
5 480 conservation of unique biodiversity (Vezina et al [2020](#)). This park is surrounded by nine GMAs
6
7 481 which provide economic and ecological buffer zones for KNP (Dietz [2021](#), Agnes [2015](#)).
8
9 482 Meanwhile, KNP has attracted significant investment in conservation initiatives. A notable
10
11 483 example is that the United Nations Development Programme launched a conservation project to
12
13 484 support the management of KNP (United Nations Development Programme [2011](#), African Parks
14
15 485 [2021](#)). KNP is the oldest and largest national park in Zambia, and the Zambia Wildlife Authority
16
17 486 (ZAWA) is solely responsible for the management of KNP (Chanda [2007](#)). This dedicated
18
19 487 management, backed by the strong support of the Zambian Government, ensures the effective
20
21 488 protection of the KNP. Hence, the prevention of deforestation in KNP is a result of the combined
22
23 489 efforts of conservation initiatives.

24
25 490 In contrast to KNP, the Kalomo region prioritizes economic development over ecological
26
27 491 conservation. The Kalomo district, known for its strong tradition of maize and livestock farming,
28
29 492 is largely acknowledged as the agricultural hub of Zambia (Moombe et al [2020](#)). The
30
31 493 implementation of measures promoting the cultivation of maize has significantly contributed to
32
33 494 the economic development of the Kalomo district (Amondo et al [2019](#)). The government
34
35 495 supported through the Fertilizer Support Program (FSP), has generated a significant increase in
36
37 496 maize production in Zambia since 2002/03, increasing the number of smallholder farmers, and
38
39 497 the maize area cultivated by smallholders also increased from about 750,000 hectares in 2002/03
40
41 498 to 1,300,000 hectares in 2010/11 (Chamberlin et al [2014](#)). The significant promotion of maize
42
43 499 production between 2000 and 2010 has led to Kalomo being one area with a major maize surplus
44
45 500 (USAID [2017](#)).

46
47 501 Despite operating under separate land management policies, KNP and Kalomo encounter
48
49 502 distinct difficulties. KNP appears to be a classic case of strict, yet repressive, conservation in
50
51 503 practice in Zambia. Its substantial restrictions on access to natural resources bear profound socio-
52
53 504 economic consequences for the adjacent GMA communities (Vezina et al [2020](#)). Communities
54
55 505 proximate to the KNP are heavily dependent on natural resources for their livelihoods but are
56
57 506 excluded from the park to the extent that most people in these communities (erroneously)
58
59 507 consider visiting the park to be illegal (Watson et al [2014](#), Namukonde and Kachali [2015](#), Milupi
60
508 et al [2021](#)). These limitations have resulted in a lack of access to food and heightened strain on

resources, which in turn undermines the effectiveness of KNP's conservation endeavours beyond the park's boundaries. In addition, the demand for land in the GMAs can rise due to ongoing population expansion, and open spaces are expected to grow in the next few years, which could undermine KNP's conservation efficacy. Indeed, evidence of settlements is already visible at the border of KNP (Dietz [2021](#)). Therefore, it is evident that a sectoral approach to conservation within KNP, that fails to consider local socio-economic needs, while conserving biodiversity within the park, is perpetuating environmental collapse beyond the park barriers. To address this issue, a more rigorous approach is needed, including awareness and education programmes to engage local communities and improve their access to the parks, as well as a more equitable share of the benefits generated by the park.

Conversely, the Kalomo district stands as a typical case where economic development is achieved at the expense of environmental sacrifice, and the uncoordinated governance on land has enlarged environmental problems. Agriculture development has contributed to livelihoods and economic growth but led to severe deforestation in Kalomo (Moombe et al [2020](#), Mbanga et al [2021](#)). In addition, Mbanga et al ([2021](#)) stated that agricultural land is expanding without proper monitoring and planning. Furthermore, forest resource management was centralized, and local communities and other stakeholders were excluded from the forest management and forest-resource-utilization systems (Wang et al [2022a](#)). This region is facing increasing pressure on the land due to uncoordinated governance (Upla et al [2022](#)), and rapid population growth, heavy reliance on agriculture for the economy, and declining soil fertility may further exacerbate this pressure. Hence, it is essential to develop more coordinated landscape management plans that harmonize local livelihood concerns with conservation targets and promote sustainable agricultural practices that enhance productivity while minimizing the negative impact on the environment. We provide more specific recommendations for improving landscape management in southern Zambia below.

5. Recommendations

5.1 Recommendation for the management of PAs

In Zambia, PAs cover a significant portion of the land (40%), with 20 national parks (64,000km²) and 36 GMAs (167,000km²) established for the primary goal of conserving

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2
3 538 biodiversity (Lindsey et al [2014](#), Hou-Jones et al [2019](#), Lecina-Diaz et al [2019](#)), with Zambian
4 539 national parks regarded as strict PAs where human settlement is not permitted (Lindsey et al
5 [2014](#)). However, the traditional approach of establishing strict PAs has been criticized for
6 540 [2014](#)). However, the traditional approach of establishing strict PAs has been criticized for
7 541 inadequately considering the needs of local communities (Mfuno [2014](#), Vasquez and Sunderland
8 542 [2023](#)), leading to conflicts over resources and increasing pressure on the surrounding landscapes
9 543 (Vezina et al [2020](#)). A sectoral approach to strict PAs to protect biodiversity and ecosystem
10 544 services that disregard local well-being needs is unlikely to be successful over the long term in
11 545 Zambia (Batáry et al [2011](#)). The effectiveness of PAs is strongly related to conservation
12 546 governance and policy frameworks, and the most positive results can be seen when Indigenous
13 547 Peoples and local communities play a central role in decision-making and have clear lines of
14 548 authority (Dehmel et al [2022](#)). Therefore, a more holistic landscape approach that integrates the
15 549 management of national parks and GMAs and considers broader landscape socio-cultural and
16 550 political-economic dynamics should be prioritized to better harmonize conservation and
17 551 development objectives.

18
19 552 To achieve the long-term conservation and sustainability of PAs, several
20 553 recommendations can be made. Firstly, involving local communities in decision-making
21 554 processes related to the management of PAs and broader land-use planning processes is crucial
22 555 to better understand their perceptions, incorporate their knowledge and needs, and ensure their
23 556 support for specific, contextually appropriate types of conservation efforts; the long-term success
24 557 of conservation is largely dependent on their support. Secondly, increasing commitments to
25 558 monitoring land-use change and strengthening cross-scale and multi-sector dialogue can
26 559 contribute to preventing the expansion of settlements closer to PAs and preserving biodiversity
27 560 and ecosystem services through clarification of rights, responsibilities, and access to information
28 561 and resources. Finally, exploration of alternative culturally appropriate livelihood strategies that
29 562 are pro-environment or sustainably used, rather than simply depleting the natural resource base
30 563 can help reduce environmental pressure and improve local well-being. Such an integrated
31 564 approach can significantly contribute towards the goals and targets of the Global Biodiversity
32 565 Framework by seeking to conserve existing PAs but also enhance ecosystem connectivity,
33 566 support restoration of surrounding degraded areas, strengthen landscape resilience, and ensure
34 567 the viability and sustainability of livelihood activities, particularly in neighbouring communities
35 568 with high natural resource dependence.

5.2 Recommendations for the management of agricultural landscapes

Zambia has suffered high rates of deforestation, driven largely by agricultural expansion (Richardson et al [2021](#)), resulting in an annual deforestation rate of 250,000 to 300,000 hectares per year (Phiri et al [2022](#)). In addition, rainfed agricultural systems are susceptible to extreme climatic events, such as more frequent, intense, and extended droughts (Black et al [2016](#), Ngoma et al [2021](#)). To address these challenges, it is necessary to pursue more sustainable and diversified agricultural production that avoids contributing to further land clearing, thereby promoting agriculture development and biodiversity conservation while responding to the impacts of climate change.

Agri-environmental management (AEM) has been identified as a crucial strategy for biodiversity conservation in cropland (Batáry et al [2011](#), Mfuno [2014](#), Batáry et al [2015](#)), with agricultural landscapes increasingly recognized as domains for conserving biodiversity and managing ES (Leakey [2012](#), Reed et al [2016](#)). Conservation agriculture (CA) is one of the significant measures of AEM (Mfuno [2014](#)), and has the potential to resolve conflicts between biodiversity conservation and economic development by increasing crop yields and diversifying crop types, thereby improving the livelihoods of farmers while reducing environmental risks (Mfuno [2014](#), FAO and UNDP [2020](#)). Therefore, CA could be promoted in agricultural landscapes in southern Zambia, and recent research suggested that there is local demand to increase the capacity for such approaches in Kalomo (Reed et al [2022](#)). Meanwhile, incorporating Indigenous and local perspectives within such processes can further strengthen integrated landscape management due to their role as holders of specific place-based social-ecological knowledge. In Zambia, the traditional knowledge of the Tonga people has contributed to the development of sustainable livelihood practices and agricultural methods, enabling them to live sustainably (Yanou et al [2023](#)). Therefore, promoting the engagement of multiple stakeholder groups, including Indigenous People and the local community, in land-use planning and natural resource management can help to generate more suitable land management solutions that satisfy the needs of humanity (food and energy production), while mitigating environmental harm (deforestation) within agricultural landscapes.

6. Conclusion

This study employed both quantitative and qualitative methods to assess the impacts of LULC changes on the ecosystem functionality of tropical landscapes in southern Zambia between 2000 and 2020. We used the most reliable and recent datasets to spatially assess LULC change, carbon storage, and the trumpeter hornbill's habitat quality. We visualized these changes on 30-meter resolution maps, providing an important resource for decision-makers and managers for future contextualized natural resource management and land-use planning interventions. Our assessment involved a comparative analysis of ES changes in two distinct landscapes: a protected area (KNP) and an agricultural landscape (Kalomo) over time. The results revealed the direct influence of policy implementation on LULC, thereby significantly affecting the functionality of landscapes that provide ES.

In summary, our study offers valuable insights into the spatial distribution of carbon storage and the habitat of the trumpeter hornbill. The practical assessment of carbon stock and habitat quality of specific species could guide how these carbon and biodiversity targets can be incorporated into national policies and implemented in climate change mitigation and adaptation (Soto-Navarro et al [2020](#), Munang et al [2013](#), Sintayehu [2018](#)). In addition, the conservation of trumpeter hornbill is important and closely tied to forest management practices in Zambia, as this species can facilitate functional connectivity of the landscapes. By assessing the trumpeter hornbill's habitat quality, our research empowers governments and stakeholders to specifically express their conservation and restoration objectives for ecosystem services in geospatial.

Our study highlights a crucial contrast: while strict enforcement measures have succeeded in conserving biodiversity habitats within KNP, a mix of pressures has led to rapid environmental degradation outside the park and across the Kalomo district. To move towards more sustainable and equitable landscape management that responds to the goals of the Global Biodiversity Framework, we suggest a need for a more holistic approach in the region that better accounts for local livelihood needs and broader political-economic and social dynamics.

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